



# Mechanism of the selective catalytic reduction of NO<sub>x</sub> with NH<sub>3</sub> over environmental-friendly iron titanate catalyst

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## ABSTRACT

The reaction mechanism of the selective catalytic reduction (SCR) of NO<sub>x</sub> with NH<sub>3</sub> over environmental-friendly iron titanate catalyst (FeTiO<sub>x</sub>) was investigated in detail. Over the iron titanate crystallite with specific Fe–O–Ti structure, both Brønsted and Lewis acid sites were present and involved in the SCR reaction. NH<sub>3</sub> mainly adsorbed on titanium sites in the form of ionic NH<sub>4</sub><sup>+</sup> and coordinated NH<sub>3</sub>, while NO<sub>x</sub> mainly adsorbed on iron sites in the form of monodentate nitrate. In a relatively low temperature range (<200 °C), the SCR process mainly followed the Langmuir–Hinshelwood (L–H) mechanism, in which the formation of monodentate nitrate from NO oxidation by O<sub>2</sub> over Fe<sup>3+</sup> was the rate-determining step. In contrast, in a relatively high temperature range (>200 °C), the SCR process mainly followed the Eley–Rideal (E–R) mechanism, in which the formation of NH<sub>2</sub>NO intermediate species following the H-abstraction of NH<sub>3</sub> by neighboring Fe<sup>3+</sup> was the rate-determining step.

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## 1. Introduction

Nitrogen oxides (NO<sub>x</sub>) producing from stationary and mobile sources, such as coal-fired power plants and diesel engines, have been a major air pollutant deteriorating the atmospheric environment. To meet more and more stringent environmental standards established worldwide for NO<sub>x</sub> emission control, selective catalytic reduction (SCR) of NO<sub>x</sub> over catalytic materials with various reducing agents (such as ammonia, urea, hydrocarbons, etc.) has been well studied. NH<sub>3</sub>-SCR is a well-proven and effective technology for the elimination of NO<sub>x</sub> from stationary sources, which has been industrially utilized since 1970s. Besides, it is also one of the most promising technologies for the removal of NO<sub>x</sub> from diesel engines. For NH<sub>3</sub>-SCR process, the most widely used catalyst system is V<sub>2</sub>O<sub>5</sub>–WO<sub>3</sub> (MoO<sub>3</sub>)/TiO<sub>2</sub>, in which the active phase V<sub>2</sub>O<sub>5</sub> is yet biologically toxic. Nowadays, many researchers are focusing on the development of new environmental-friendly NH<sub>3</sub>-SCR catalysts to substitute the conventional vanadium-based catalyst, which also shows some other problems in practical use, such as narrow operation temperature window, low N<sub>2</sub> selectivity and

high conversion of SO<sub>2</sub> to SO<sub>3</sub> at high temperatures. In our previous works [1–4], we have successfully developed a novel and environmental-friendly iron titanate catalyst (FeTiO<sub>x</sub>) with specific Fe–O–Ti structure, showing excellent NH<sub>3</sub>-SCR activity, N<sub>2</sub> selectivity and H<sub>2</sub>O/SO<sub>2</sub> durability in the medium temperature range (200–400 °C). However, the SCR mechanism over this catalyst is still unclear.

Even for V<sub>2</sub>O<sub>5</sub>–WO<sub>3</sub> (MoO<sub>3</sub>)/TiO<sub>2</sub> catalyst, which has been used for decades, several debates in mechanistic aspects still remain, such as the intrinsic nature of active sites [5] and reactive adsorbed NH<sub>3</sub> species (bound to Brønsted or Lewis acid sites) involved in the SCR reaction [6]. Similarly, different SCR mechanisms have been proposed over Fe-based catalysts, including Langmuir–Hinshelwood (L–H) mechanism and Eley–Rideal (E–R) mechanism. Over Fe<sup>3+</sup>–TiO<sub>2</sub>–PILC catalyst prepared by Long and Yang [6], they proposed that both ionic NH<sub>4</sub><sup>+</sup> and coordinated NH<sub>3</sub> could react with adsorbed NO<sub>2</sub> to form intermediate species, which could further react with NO to form N<sub>2</sub> and H<sub>2</sub>O, following an L–H mechanism. Whereas, over Fe–ZSM-5 catalyst prepared by Long and Yang [7], it was proposed that only ionic NH<sub>4</sub><sup>+</sup> contributed to the SCR reaction, following a similar L–H mechanism. Contrastively, Apostolescu et al. pointed out that over Fe<sub>2</sub>O<sub>3</sub>–WO<sub>3</sub>/ZrO<sub>2</sub> catalyst the SCR reaction mainly followed an E–R mechanism, in which only Lewis acid sites were involved in NH<sub>3</sub> activation and NO reduction process [8]. Therefore, the SCR mechanisms proposed by researchers over different catalysts are usually diverse. It is necessary to get a comprehensive understanding of the reac-

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tion mechanism over FeTiO<sub>x</sub> catalyst synthesized by our group for further improving its low temperature SCR activity and H<sub>2</sub>O/SO<sub>2</sub> durability for industrial application.

In this article, temperature programmed desorption (TPD), temperature programmed surface reaction (TPSR), transient response (TR) and *in situ* diffuse reflectance infrared Fourier transform spectroscopy (*in situ* DRIFTS) will be used for NH<sub>3</sub>-SCR mechanism study over FeTiO<sub>x</sub> catalyst. The reactive surface species including adsorbed NH<sub>3</sub> and NO<sub>x</sub> species will be clearly clarified. Two different mechanisms in the relatively low and high temperature ranges will be proposed accordingly, which can supply theoretical guidance for the catalyst redesign and thus the activity improvement.

## 2. Experimental

### 2.1. Catalyst preparation, SCR activity and characterizations

FeTiO<sub>x</sub> catalyst and Fe<sub>2</sub>O<sub>3</sub>, TiO<sub>2</sub> reference samples were prepared by co-precipitation method using Fe(NO<sub>3</sub>)<sub>3</sub>·9H<sub>2</sub>O and Ti(SO<sub>4</sub>)<sub>2</sub> as precursors [2]. During the preparation process of FeTiO<sub>x</sub> catalyst, Ti(SO<sub>4</sub>)<sub>2</sub> and Fe(NO<sub>3</sub>)<sub>3</sub>·9H<sub>2</sub>O were firstly dissolved together with distilled water with the molar ratio of Fe:Ti = 1:1. Afterwards, standard NH<sub>3</sub>-H<sub>2</sub>O (25 wt.%) was used as precipitator until the pH rose to 10 when the Fe and Ti ions were completely co-precipitated. Without aging, the precipitate was filtrated and washed, followed by desiccation at 100 °C for 12 h and calcination at 400 °C for 6 h in air condition. The calcined sample was crushed and sieved to 20–40 mesh for activity test and TPD, TPSR and TR experiments, and above 120 mesh for *in situ* DRIFTS experiments. As for reference samples, Fe<sub>2</sub>O<sub>3</sub> and TiO<sub>2</sub> were self-prepared by precipitation method using Fe(NO<sub>3</sub>)<sub>3</sub>·9H<sub>2</sub>O and Ti(SO<sub>4</sub>)<sub>2</sub> as precursors, respectively. The desiccation, calcination and sieving procedures were controlled exactly the same as those of FeTiO<sub>x</sub> catalyst.

Our previous study [2] showed that FeTiO<sub>x</sub> catalyst showed a good NH<sub>3</sub>-SCR activity in the temperature range of 200–350 °C with the NO<sub>x</sub> conversion above 90%, which was 50–150 °C lower than those of Fe/ZSM-5 [9], Fe-TiO<sub>2</sub>-PILC [10], Fe<sub>2</sub>O<sub>3</sub>/WO<sub>3</sub>/ZrO<sub>2</sub> [8] and Fe/HBEA catalysts [11]. Based on the calculation results of turn-over frequency (TOF), the intrinsic activity of FeTiO<sub>x</sub> catalyst was also comparable to that of the state-of-the-art SCR catalyst V<sub>2</sub>O<sub>5</sub>-WO<sub>3</sub>/TiO<sub>2</sub> or commercial Fe-exchanged BEA catalyst [11], but still much lower than that of Fe/ZSM-5 or Cu/ZSM-5 reported by other researchers [12,13] probably because this FeTiO<sub>x</sub> catalyst possessed a majority of ineffective iron species in the bulk phase. According to the XRD, UV-Vis DRS, Raman spectroscopy and XAFS results, the FeTiO<sub>x</sub> catalyst was mainly in the form of iron titanate crystallite with specific Fe-O-Ti structure, while the reference samples Fe<sub>2</sub>O<sub>3</sub> and TiO<sub>2</sub> were mainly present as well-crystallized hematite and anatase, respectively [2]. The BET surface areas of FeTiO<sub>x</sub>, Fe<sub>2</sub>O<sub>3</sub> and TiO<sub>2</sub> obtained from N<sub>2</sub> physisorption results were 245.3, 42.5 and 103.5 m<sup>2</sup>/g, respectively, and the BJH desorption pore volumes were 0.52, 0.21 and 0.20 cm<sup>3</sup> g<sup>-1</sup>, respectively [2].

### 2.2. TPD, TPSR and TR experiments

TPD, TPSR and TR experiments were performed over 200 mg samples in a fixed-bed quartz tube reactor using a quadrupole mass spectrometer (HPR20, Hiden Analytical Ltd., Amplifier Head HAL 301) to record the reactant signals. The recorded signals were as follows: *m/z* = 16 (NH<sub>2</sub>) or 15 (NH) to identify NH<sub>3</sub> in the absence or presence of O<sub>2</sub>; *m/z* = 30 to identify NO or NO<sub>2</sub>; *m/z* = 18, 28, 32 and 44 to identify H<sub>2</sub>O, N<sub>2</sub>, O<sub>2</sub> and N<sub>2</sub>O, respectively. Due to the very small relevant abundance of H<sub>2</sub>O signal with *m/z* = 16 (*ca.* 0.9% comparing with *m/z* = 18 obtained from NIST Mass Spectrom-

etry Data Center), we can distinguish NH<sub>3</sub> and H<sub>2</sub>O species in mass spectrometer clearly.

Prior to the TPD, TPSR and TR experiments, the samples were pretreated in a flow of 20 vol% O<sub>2</sub>/He (30 ml/min) at 300 °C for 30 min. After cooling down to room temperature, the O<sub>2</sub>/He flow was switched to a flow of 2500 ppm NH<sub>3</sub>/Ar or 2500 ppm NO + 10 vol% O<sub>2</sub>/Ar (30 ml/min) for 1 h. Then the samples were purged with Ar for another 1 h until the mass spectrometer signals were stabilized. For TPD experiments, the temperature was directly raised linearly to 500 °C at a rate of 10 °C/min in a flow of Ar.

For TPSR experiments between adsorbed NH<sub>3</sub> species with NO or NO + O<sub>2</sub>, subsequent to the NH<sub>3</sub> adsorption and Ar purge, the Ar flow was switched to different flows of 500 ppm NO/Ar or 500 ppm NO + 5 vol% O<sub>2</sub>/Ar for 1 h until the concentrations of reactants were stabilized. Then the temperature was raised linearly to 500 °C at a rate of 10 °C/min. On the contrary, for TPSR experiment between adsorbed NO<sub>x</sub> species with NH<sub>3</sub>, subsequent to the NO + O<sub>2</sub> adsorption and Ar purge, the Ar flow was switched to a flow of 500 ppm NH<sub>3</sub>/Ar for 1 h until the concentration of NH<sub>3</sub> was stabilized. Then the temperature was also raised linearly to 500 °C at a rate of 10 °C/min.

For TR experiments, subsequent to the NH<sub>3</sub> or NO + O<sub>2</sub> adsorption and Ar purge at room temperature, the temperature was raised to 200 °C or 150 °C and held on for 1 h until the partial desorption of surface species at this temperature was complete. Then NO + O<sub>2</sub> or NH<sub>3</sub> was turned on and off repeatedly to observe the pulse reaction between the gaseous reactants with adsorbed species. For TR experiments at 150 °C, Amplifier Head HAL 201 equipped on HPR20 mass spectrometer was used.

### 2.3. *In situ* DRIFTS

The *in situ* DRIFTS experiments were performed on an FTIR spectrometer (Nicolet Nexus 670) equipped with an MCT/A detector cooled by liquid nitrogen. An *in situ* reactor cell with ZnSe window (Nexus Smart Collector) connected to an adsorption/purging gas control system was used. The reaction temperature was controlled precisely by an Omega programmable temperature controller. Prior to each *in situ* DRIFTS experiment, the sample was pretreated at 400 °C in a flow of 20 vol% O<sub>2</sub>/N<sub>2</sub> for 30 min and then cooled down to the desired reaction temperature (200 °C). The spectrum of FeTiO<sub>x</sub> catalyst at 200 °C was collected in flowing N<sub>2</sub> atmosphere and set as background, which was automatically subtracted from the final spectrum. The total flow rate of the feeding gas was kept at 300 ml/min and all spectra were recorded by accumulating 100 scans with a resolution of 4 cm<sup>-1</sup>. The reaction conditions were controlled as follows: 500 ppm NO, 500 ppm NH<sub>3</sub>, 5 vol% O<sub>2</sub> and N<sub>2</sub> balance.

## 3. Results

### 3.1. NH<sub>3</sub>-TPD and NO<sub>x</sub>-TPD experiments

The adsorption/desorption capacity of NH<sub>3</sub> and NO<sub>x</sub> over FeTiO<sub>x</sub> catalyst was investigated by TPD methods using TiO<sub>2</sub> and Fe<sub>2</sub>O<sub>3</sub> as reference samples (Fig. 1). As the NH<sub>3</sub>-TPD results shown in Fig. 1A, all of the samples had three NH<sub>3</sub> desorption peaks from 30 to 500 °C. The peaks below 100 °C can be ascribed to physisorbed NH<sub>3</sub>, and the peaks located at 100–120 °C can be ascribed to NH<sub>4</sub><sup>+</sup> bound to weak Brønsted acid sites (some surface hydroxyls) [14,15]. For FeTiO<sub>x</sub> and TiO<sub>2</sub>, the desorption peaks above 200 °C can be ascribed to NH<sub>4</sub><sup>+</sup> bound to strong Brønsted acid sites (some surface hydroxyls with enhanced acidity induced by sulfate species and hydroxyls directly linked to sulfate species) and coordinated NH<sub>3</sub> bound

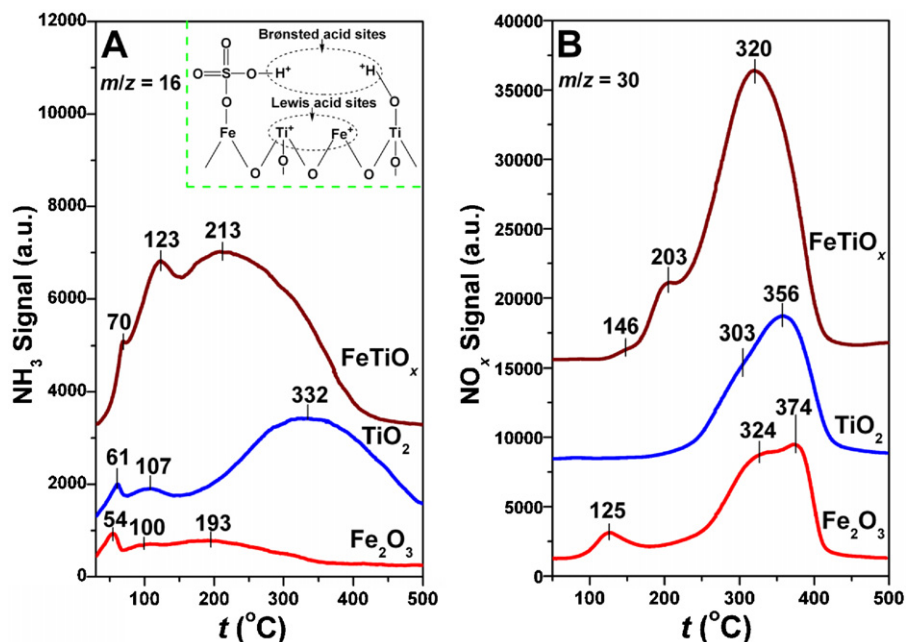


Fig. 1. (A)  $\text{NH}_3$ -TPD and (B)  $\text{NO}_x$ -TPD profiles over  $\text{FeTiO}_x$ ,  $\text{TiO}_2$  and  $\text{Fe}_2\text{O}_3$ .

to Lewis acid sites [16]. For  $\text{Fe}_2\text{O}_3$ , a lower and broader desorption peak was observed centered at 193  $^\circ\text{C}$ , which can be only attributed to  $\text{NH}_3$  bound to Lewis acid sites because no sulfate species thus no strong Brønsted acid sites existed on this sample. In the whole temperature range,  $\text{FeTiO}_x$  catalyst showed the maximum  $\text{NH}_3$  adsorption and consequently desorption amount. The calculated  $\text{NH}_3$  adsorption amounts on these three samples decreased in the following sequence:  $\text{FeTiO}_x$  (68.6  $\mu\text{mol/g}$ ) >  $\text{TiO}_2$  (40.0  $\mu\text{mol/g}$ )  $\gg$   $\text{Fe}_2\text{O}_3$  (8.7  $\mu\text{mol/g}$ ).

The interaction between iron and titanium species in  $\text{FeTiO}_x$  resulted in larger amount of Brønsted acid sites per weight than those on pure  $\text{TiO}_2$  and  $\text{Fe}_2\text{O}_3$ , which was beneficial to the SCR reaction. This result is in well accordance with the *in situ* DRIFTS of  $\text{NH}_3$  adsorption over these three samples (Fig. S1 in Supporting information). After  $\text{NH}_3$  adsorption at 30  $^\circ\text{C}$ , the bands of  $\text{NH}_4^+$  adsorbed on  $\text{FeTiO}_x$  showed a much higher intensity than those on  $\text{TiO}_2$ , and the amount of  $\text{NH}_4^+$  species detected on  $\text{Fe}_2\text{O}_3$  was rather small. After normalization by BET surface areas, the  $\text{NH}_3$  adsorption amounts on these three samples decreased in the following sequence:  $\text{TiO}_2$  (0.39  $\mu\text{mol/m}^2$ ) >  $\text{FeTiO}_x$  (0.28  $\mu\text{mol/m}^2$ ) >  $\text{Fe}_2\text{O}_3$  (0.20  $\mu\text{mol/m}^2$ ), indicating that Ti species showed higher  $\text{NH}_3$  adsorption ability than Fe species. Based on these results, a model for Brønsted and Lewis acid sites on  $\text{FeTiO}_x$  catalyst was proposed, as shown in the inset of Fig. 1A. The Brønsted acid sites were mainly composed of acidic hydroxyls linked to  $\text{Ti}^{4+}$  and sulfate species, while the Lewis acid sites were mainly composed of unsaturated  $\text{Ti}^{4+}$  and  $\text{Fe}^{3+}$ . From  $\text{NH}_3$  desorption amounts over  $\text{FeTiO}_x$ ,  $\text{TiO}_2$  and  $\text{Fe}_2\text{O}_3$ , it is concluded that  $\text{NH}_3$  mainly adsorbed on titanium sites, but not iron sites of  $\text{FeTiO}_x$  catalyst in the form of ionic  $\text{NH}_4^+$  and coordinated  $\text{NH}_3$ , which could further participate in the SCR reaction.

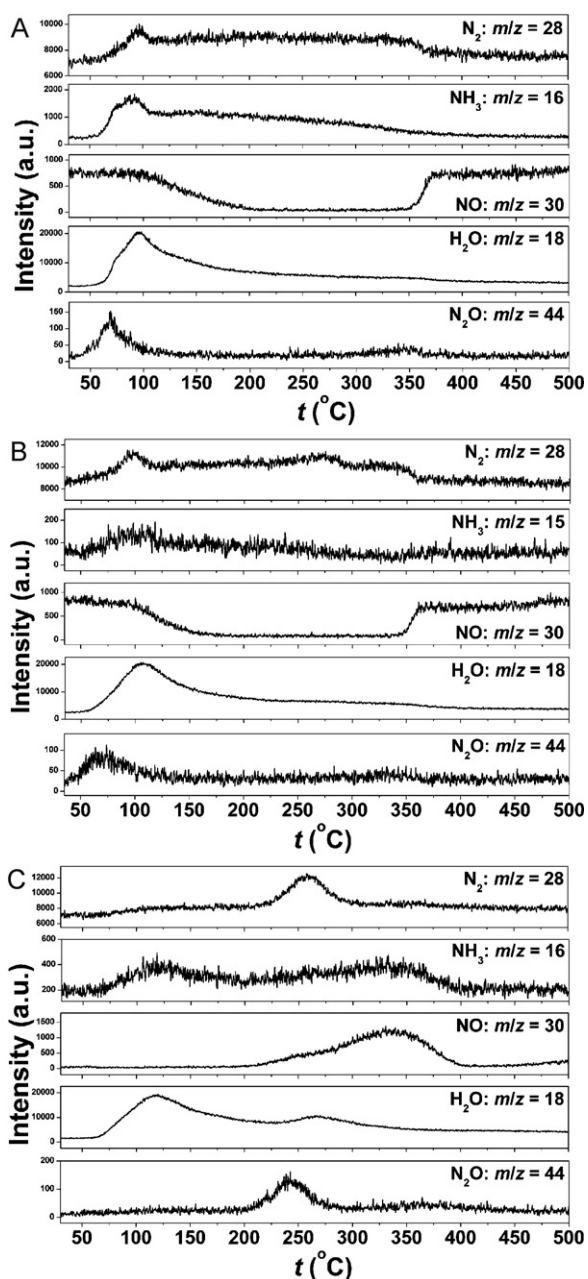
Fig. 1B shows the  $\text{NO}_x$ -TPD profiles over three samples, among which  $\text{FeTiO}_x$  catalyst had the largest  $\text{NO}_x$  desorption amount. It was reported that [17] physisorbed  $\text{NO}_x$  usually desorbed from catalyst surface below 100  $^\circ\text{C}$ , and according to this, no physisorbed  $\text{NO}_x$  was detected over all samples in this study. Both  $\text{FeTiO}_x$  and  $\text{Fe}_2\text{O}_3$  had three desorption peaks, whereas  $\text{TiO}_2$  only showed two. Comparing with the *in situ* DRIFTS of  $\text{NO}_x$  desorption over  $\text{FeTiO}_x$  and  $\text{Fe}_2\text{O}_3$  (Fig. S2 in Supporting information), the desorption peaks below 150  $^\circ\text{C}$  can be ascribed to the decomposition of monodentate nitrate. The desorption peaks centered at 203  $^\circ\text{C}$  for  $\text{FeTiO}_x$  and 324  $^\circ\text{C}$  for  $\text{Fe}_2\text{O}_3$  can be ascribed to the decomposition of bridging nitrate with different thermal stability. The desorption peaks in the high temperature range centered at 320  $^\circ\text{C}$  for  $\text{FeTiO}_x$  and 374  $^\circ\text{C}$  for  $\text{Fe}_2\text{O}_3$  are assigned to the decomposition of bidentate nitrate. For  $\text{TiO}_2$ , the two desorption peaks at 303 and 356  $^\circ\text{C}$  can be assigned to the decomposition of bridging nitrate and bidentate nitrate, respectively.

From  $\text{NO}_x$  desorption amounts over these three samples, we can see that  $\text{NO}_x$  could adsorb on both iron sites and titanium sites over  $\text{FeTiO}_x$  catalyst. However, previous study [5] showed that  $\text{NO}_x$  could not adsorb on  $\text{NH}_3$  pre-covered  $\text{Ti}^{4+}$  sites due to the greater basicity of  $\text{NH}_3$ . Considering the  $\text{NH}_3$ -TPD results which already showed that  $\text{NH}_3$  mainly adsorbed on  $\text{Ti}^{4+}$  sites, it is deduced that under the SCR condition  $\text{NO}_x$  mainly adsorbed on  $\text{Fe}^{3+}$  sites in the form of monodentate nitrate or bidentate nitrate, rather than bridging nitrate between  $\text{Fe}^{3+}$  and  $\text{Ti}^{4+}$  sites because of the blockage of  $\text{Ti}^{4+}$  sites in Fe–O–Ti structure by adsorbed  $\text{NH}_3$  species. Subsequent TPRS and *in situ* DRIFTS study will reveal what kinds of nitrate species can finally be transformed into  $\text{N}_2$  in the SCR condition.

3.2. TPRS experiments

Fig. 2A shows the results of TPRS between NO and pre-adsorbed  $\text{NH}_3$  species over  $\text{FeTiO}_x$ . Below 100  $^\circ\text{C}$ ,  $\text{NH}_3$  and  $\text{H}_2\text{O}$  desorption was significant and NO concentration showed no decrease, which meant that the SCR reaction did not occur. The  $\text{N}_2$  and  $\text{N}_2\text{O}$  formation below 100  $^\circ\text{C}$  might be attributed to the decomposition of some surface nitrogenous species, which were formed through the reaction between  $\text{NO}_x$  and adsorbed  $\text{NH}_3$  species in presence of  $\text{H}_2\text{O}$  at room temperature [17]. Above 100  $^\circ\text{C}$ , NO concentration had a steady decrease, implying the occurrence of SCR reaction, and the minimum NO concentration was obtained from 200 to 350  $^\circ\text{C}$ . In this temperature range, both  $\text{N}_2$  and  $\text{H}_2\text{O}$  had a broad band above baseline and no  $\text{N}_2\text{O}$  formation was observed, suggesting that the selectivity to  $\text{N}_2$  was rather high. Above 350  $^\circ\text{C}$ , the pre-stored  $\text{NH}_3$  species on the catalyst surface seemed to be totally consumed, resulting in the increase of NO and simultaneous decrease of  $\text{N}_2$  in gas phase. This TPRS result showed that NO could react with





**Fig. 2.** TPRS profiles of (A) NO with adsorbed  $\text{NH}_3$  species, (B)  $\text{NO} + \text{O}_2$  with adsorbed  $\text{NH}_3$  species, and (C)  $\text{NH}_3$  with adsorbed  $\text{NO}_x$  species over  $\text{FeTiO}_x$ .

adsorbed  $\text{NH}_3$  species in the absence of  $\text{O}_2$  to form  $\text{N}_2$  and  $\text{H}_2\text{O}$ , and the lattice oxygen in  $\text{FeTiO}_x$  catalyst might be involved in the SCR reaction, which is in well accordance with our previous conclusion [2].

To further investigate the effect of  $\text{O}_2$  in the SCR reaction, TPRS between  $\text{NO} + \text{O}_2$  and pre-adsorbed  $\text{NH}_3$  species was also conducted (Fig. 2B). In contrast to that with NO only in Fig. 2A, NO was consumed more rapidly to the minimum concentration at  $150^\circ\text{C}$  in the presence of  $\text{O}_2$ . An enlarged operation temperature window was thus obtained. This result showed that  $\text{O}_2$  played an important role in the promotion of SCR reaction, especially in the low temperature range where the NO oxidation by  $\text{O}_2$  was significant. The TPRS results with and without  $\text{O}_2$  also implied in an opposite side that NO could participate in the SCR reaction in gaseous or weakly adsorbed phase at relatively high temperatures ( $>200^\circ\text{C}$ ) without undergoing the oxidation step by gaseous  $\text{O}_2$ .

Moreover, TPRS experiment between  $\text{NH}_3$  and pre-adsorbed  $\text{NO}_x$  species was conducted (Fig. 2C). With the increase of reaction temperature, both  $\text{NH}_3$  and  $\text{H}_2\text{O}$  firstly had a desorption band around  $150^\circ\text{C}$ , and then  $\text{NH}_3$  had an obvious consumption from 200 to  $250^\circ\text{C}$ . This  $\text{NH}_3$  consumption was due to the occurrence of SCR reaction because both  $\text{N}_2$  and  $\text{H}_2\text{O}$  had an evident formation peak centered at ca.  $250^\circ\text{C}$ . A small amount of  $\text{N}_2\text{O}$  was also detected in this temperature range, implying that the pre-adsorbed nitrogenous species did not totally convert into  $\text{N}_2$  through the reaction with  $\text{NH}_3$ . Afterwards, a slight increase of NO was observed at about  $250^\circ\text{C}$ , which was caused by the decomposition of some bridging nitrate. Above  $300^\circ\text{C}$ , a large amount of  $\text{NH}_3$  and NO desorbed from the catalyst surface, and both of the  $\text{N}_2$  and  $\text{H}_2\text{O}$  concentrations returned to the baselines. Comparing with the  $\text{NO}_x$ -TPD results, this  $\text{NO}_x$  desorption peak was mainly caused by the decomposition of bidentate nitrate which had the highest thermal stability over  $\text{FeTiO}_x$  catalyst. This result showed that bidentate nitrate was not reactive in the SCR reaction, just as the spectator species. The same phenomenon was also observed in previous studies [18–21], in which bidentate nitrate was not or less reactive in the SCR reaction towards various reducing agents ( $\text{NH}_3$ ,  $\text{C}_3\text{H}_6$  or  $\text{C}_3\text{H}_8$ ). This conclusion will also be verified in the following *in situ* DRIFTS study.

### 3.3. TR experiments

To further elucidate the SCR reaction pathways over  $\text{FeTiO}_x$  catalyst, we also conducted TR experiments at fixed temperatures ( $200^\circ\text{C}$  and lower  $150^\circ\text{C}$ ). Fig. 3A shows the results of  $\text{NH}_3$  pulse reaction with pre-adsorbed  $\text{NO}_x$  species at  $200^\circ\text{C}$ , in which the arrowheads indicate the on-off of  $\text{NH}_3$ . After the first  $\text{NH}_3$  pulse, only a small amount of  $\text{N}_2$  was produced due to the reaction between  $\text{NH}_3$  and surface nitrate species.  $\text{H}_2\text{O}$  formation was not observed because its amount was too small to be detected. After the second  $\text{NH}_3$  pulse, neither  $\text{N}_2$  nor  $\text{H}_2\text{O}$  was detected, implying that the reactive nitrate species was completely consumed in the first pulse. During the whole process, the concentration of  $\text{NH}_3$  in gas phase did not change due to its strong adsorption onto the catalyst surface. These results show that at this temperature ( $200^\circ\text{C}$ ), some nitrate species indeed could react with  $\text{NH}_3$  species to form  $\text{N}_2$  following the L–H mechanism, but the amount was rather small under the reaction condition we set. However, in the real SCR condition at relatively low temperatures ( $<200^\circ\text{C}$ ), continuous formation of reactive nitrate species on iron sites in the presence of gaseous  $\text{NO} + \text{O}_2$  could occur due to its rapid reduction by adsorbed  $\text{NH}_3$  species. In order to further confirm this short conclusion, we also carried out a similar TR experiment at  $150^\circ\text{C}$ , at which more reactive nitrate species could form on the catalyst surface. As shown in Fig. 3C, more  $\text{N}_2$  resulting from the surface reaction between reactive nitrate species and adsorbed  $\text{NH}_3$  species was clearly observed after the five  $\text{NH}_3$  pulses. These results indicate that the L–H reaction pathway over  $\text{FeTiO}_x$  catalyst indeed contributed to the overall SCR reaction, especially at temperatures below  $200^\circ\text{C}$ .

For comparison, we also carried out TR experiment of  $\text{NO} + \text{O}_2$  pulse reaction with pre-adsorbed  $\text{NH}_3$  species at  $200^\circ\text{C}$  (Fig. 3B). The arrowheads indicate the on-off of  $\text{NO} + \text{O}_2$  simultaneously. Being different from the results in Fig. 3A, in the total six  $\text{NO} + \text{O}_2$  pulses, both  $\text{N}_2$  and  $\text{H}_2\text{O}$  immediately formed, implying that the  $\text{NO}_x$  in gas phase or in weakly adsorbed state could react with adsorbed  $\text{NH}_3$  species rapidly, following an E–R mechanism. After the six pulses,  $\text{NO} + \text{O}_2$  was kept open all the time to consume the residual adsorbed  $\text{NH}_3$  species. After nearly 130 min, a breakthrough curve of NO with a steady increase in concentration was obtained. At the same time, both  $\text{N}_2$  and  $\text{H}_2\text{O}$  had a steady decrease. As the similar TR experiment at  $150^\circ\text{C}$  shown in Fig. 3D, it only took ca. 70 min to obtain the breakthrough curve of NO, indicating that the E–R reaction pathway was relatively weakened to a certain

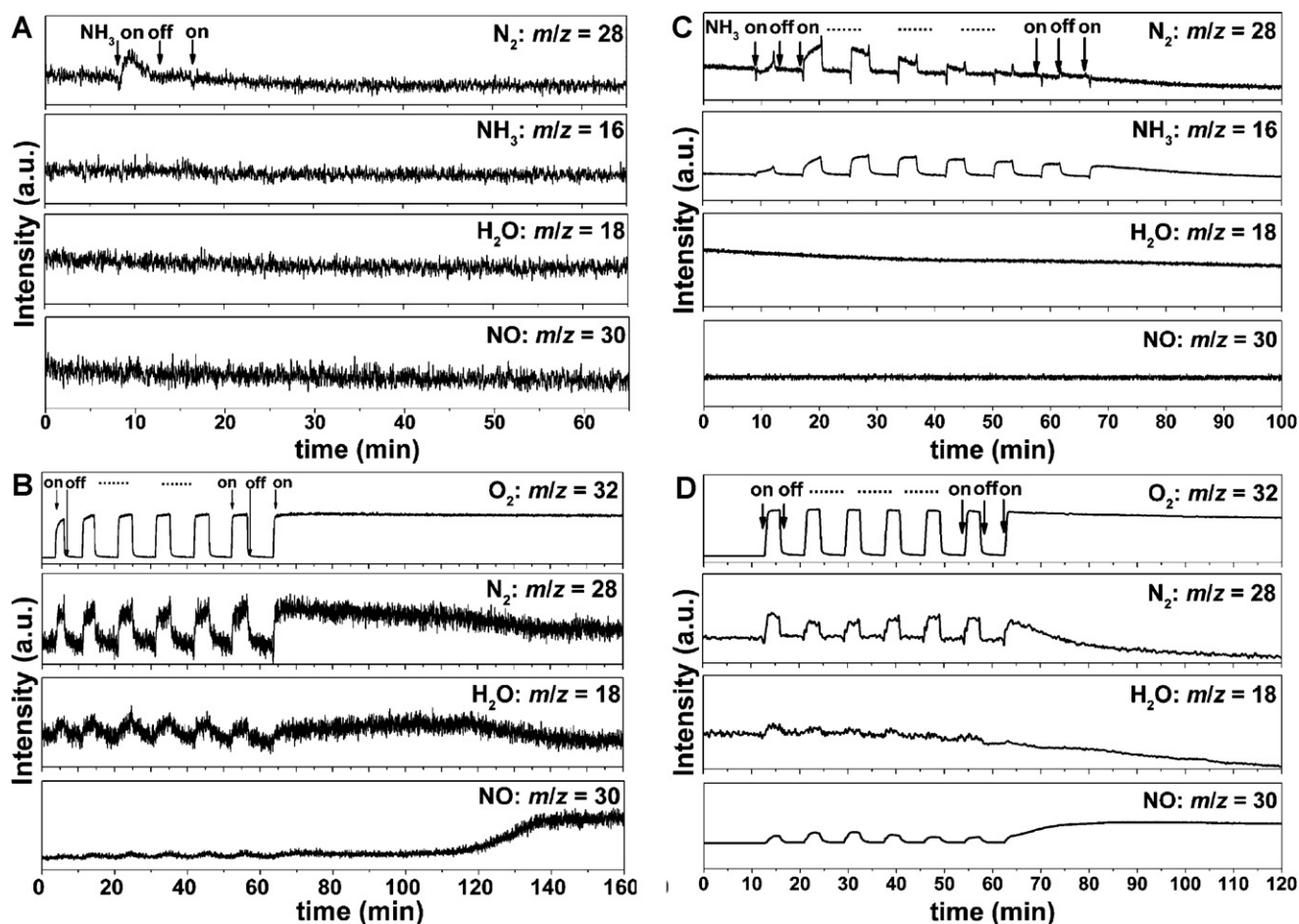


Fig. 3. TR profiles of  $\text{NH}_3$  pulse reaction with adsorbed  $\text{NO}_x$  species at (A) 200 °C and (C) 150 °C, and  $\text{NO} + \text{O}_2$  pulse reaction with adsorbed  $\text{NH}_3$  species at (B) 200 °C and (D) 150 °C over  $\text{FeTiO}_x$  catalyst.

extent at lower reaction temperatures below 200 °C. Summarizing the TR results, it is concluded that both of the reaction pathways including L–H mechanism and E–R mechanism were involved in the SCR process, as further confirmed by the following *in situ* DRIFTS results.

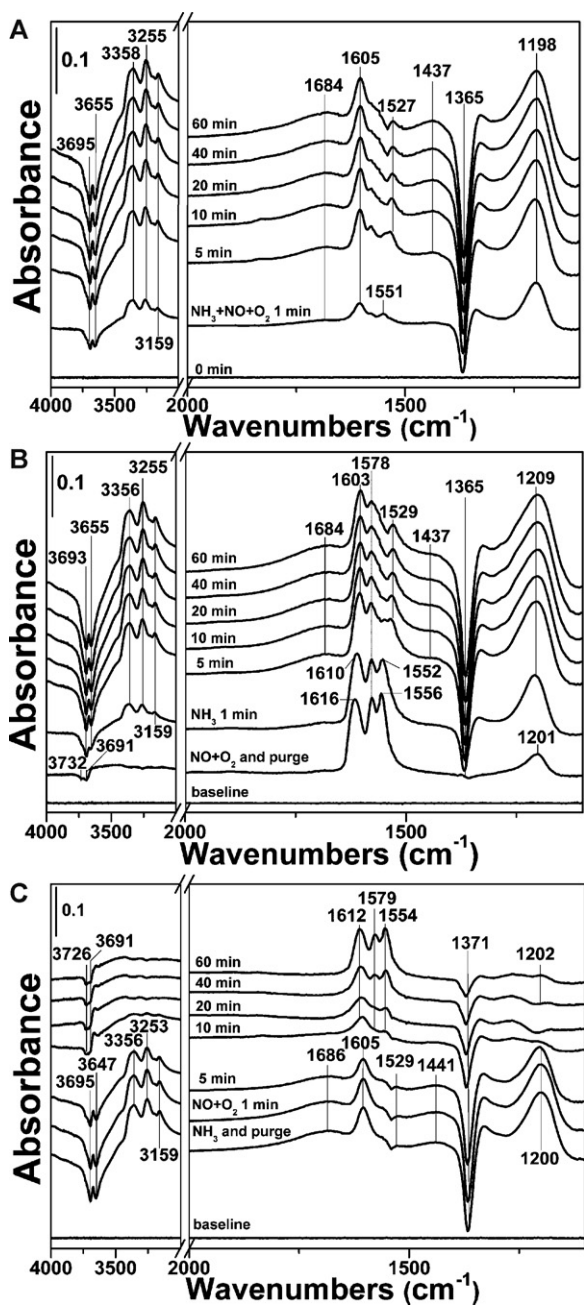
### 3.4. *In situ* DRIFTS experiments

Fig. 4A shows the *in situ* DRIFTS of reaction between  $\text{NH}_3$  and pre-adsorbed  $\text{NO}_x$  species over  $\text{FeTiO}_x$ . After  $\text{NO} + \text{O}_2$  adsorption and  $\text{N}_2$  purge, the catalyst surface was mainly covered by nitrate species. Two hydroxyl consumption bands at 3732 and 3691  $\text{cm}^{-1}$  due to the interaction between basic OH and  $\text{NO}_x$  (mainly  $\text{NO}_2$ ) [22] were observed. Bands at 1556 ( $\nu_3$  high) and 1578 ( $\nu_3$  high)  $\text{cm}^{-1}$  can be ascribed to monodentate nitrate and bidentate nitrate, respectively [18,20,23]. The assignments of the band near 1616  $\text{cm}^{-1}$  were diverse in previous studies by other researchers [18,20,24,25]. Long and Yang [6] assigned this band to adsorbed  $\text{NO}_2$  species (nitro or adsorbed  $\text{NO}_2$  molecule) due to its different thermal stability from those of other nitrate species. However, in this study we found that both of the bands at 1616 ( $\nu_3$  high) and 1201 ( $\nu_3$  low)  $\text{cm}^{-1}$  should be ascribed to bridging nitrate [23,26], because these two bands showed the same thermal stability in the *in situ* DRIFTS results of  $\text{NO}_x$  desorption (Fig. S2 in Supporting information) and same reactivity in the following reaction process.

After  $\text{NH}_3$  introduction for 1 min, the bands of monodentate nitrate and bridging nitrate immediately moved to lower wave numbers (1552 and 1610  $\text{cm}^{-1}$ ), implying the decrease of surface

coverage because of the reduction by  $\text{NH}_3$ . At the same time,  $\text{NH}_3$  strongly adsorbed onto the surface and the band at 1209  $\text{cm}^{-1}$  [27] attributed to coordinated  $\text{NH}_3$  ( $\delta_s$ ) bound to Lewis acid sites might have overlapped with the shifted band of bridging nitrate at low wave numbers (1201  $\text{cm}^{-1}$ ). After 5 min, both of the monodentate nitrate and bridging nitrate disappeared and only bidentate nitrate at 1578  $\text{cm}^{-1}$  remained on the surface, implying that bidentate nitrate was indeed inactive in the SCR reaction. Meanwhile, the IR bands attributed to coordinated  $\text{NH}_3$  ( $\delta_{as}$  at 1603  $\text{cm}^{-1}$  and  $\delta_s$  at 1209  $\text{cm}^{-1}$ ) continued growing, and the bands in the region of 3400–3100  $\text{cm}^{-1}$  attributed to N–H stretching vibration in coordinated  $\text{NH}_3$  were also observed [28]. Noticeably, a small band at 1529  $\text{cm}^{-1}$  ascribed to  $\text{NH}_2$  (scissoring vibration mode) from H-abstraction of coordinated  $\text{NH}_3$  also occurred [26,29], which was an important intermediate species to react with NO to produce nitrosamine ( $\text{NH}_2\text{NO}$ ) and then to decompose into  $\text{N}_2$  and  $\text{H}_2\text{O}$ .  $\text{NH}_4^+$  bound to Brønsted acid sites ( $\delta_s$  at 1684  $\text{cm}^{-1}$  and  $\delta_{as}$  at 1437  $\text{cm}^{-1}$ ) was also formed [30]. Interestingly, an intense negative band at 1364  $\text{cm}^{-1}$  ( $\nu_{as\text{OSO}}$ ) showed up, which is caused by the reaction between  $\text{NH}_3$  and sulfate species to form  $\text{NH}_4^+$ . This negative band proves the existence of residual sulfate species on catalyst surface, and it can also be used as an indicator for the existence of  $\text{NH}_4^+$ . Comparing with the spectrum after  $\text{NO} + \text{O}_2$  adsorption and  $\text{N}_2$  purge, the two hydroxyl consumption bands moved to 3693 and 3655  $\text{cm}^{-1}$ , which proved again that the adsorption sites of  $\text{NH}_3$  on  $\text{FeTiO}_x$  catalyst were different from those of  $\text{NO}_x$ .

Fig. 4B shows the *in situ* DRIFTS of reaction between  $\text{NO} + \text{O}_2$  and pre-adsorbed  $\text{NH}_3$  species. After  $\text{NH}_3$  adsorption and  $\text{N}_2$  purge, the



**Fig. 4.** *In situ* DRIFTS of reaction between (A)  $\text{NH}_3$  and adsorbed  $\text{NO}_x$  species, (B)  $\text{NO} + \text{O}_2$  and adsorbed  $\text{NH}_3$  species, and (C)  $\text{NH}_3 + \text{NO} + \text{O}_2$  (SCR condition) over  $\text{FeTiO}_x$  at  $200^\circ\text{C}$ .

bands belonging to coordinated  $\text{NH}_3$  ( $\delta_{\text{as}}$  at  $1605\text{ cm}^{-1}$  and  $\delta_{\text{s}}$  at  $1200\text{ cm}^{-1}$ ) and ionic  $\text{NH}_4^+$  ( $\delta_{\text{s}}$  at  $1686\text{ cm}^{-1}$  and  $\delta_{\text{as}}$  at  $1441\text{ cm}^{-1}$ ) appeared. Negative band at  $1371\text{ cm}^{-1}$ ,  $\text{NH}_2$  band at  $1529\text{ cm}^{-1}$ , N–H stretching vibration in the region of  $3400\text{--}3100\text{ cm}^{-1}$  together with two hydroxyl consumption bands at  $3695$  and  $3647\text{ cm}^{-1}$  also showed up.

After  $\text{NO} + \text{O}_2$  introduction for 5 min, no formation of nitrate species was observed, whereas the band intensity of coordinated  $\text{NH}_3$  had an obvious decrease. A slight decrease of band intensity of ionic  $\text{NH}_4^+$  was also observed, which was associated with the partial recovery of hydroxyl consumption bands and negative band at  $1371\text{ cm}^{-1}$ . This result implied that under this reaction condition both of coordinated  $\text{NH}_3$  and ionic  $\text{NH}_4^+$  could react with  $\text{NO} + \text{O}_2$ ,

and the reaction between coordinated  $\text{NH}_3$  and  $\text{NO}$  dominated following an E–R mechanism. After 10 min, coordinated  $\text{NH}_3$  species were consumed completely and nitrate species began to form on the surface. At the same time, ionic  $\text{NH}_4^+$  continued to be consumed with a faster speed that could be judged from the recovery of the negative band at  $1371\text{ cm}^{-1}$ . This meant that ionic  $\text{NH}_4^+$  favored the reaction with nitrate species following an L–H mechanism. After 60 min, the surface was mainly covered by nitrate species which contained bridging nitrate at  $1612$  and  $1202\text{ cm}^{-1}$ , bidentate nitrate at  $1579\text{ cm}^{-1}$  and monodentate nitrate at  $1554\text{ cm}^{-1}$ . The increase of bridging nitrate associating with the decrease of adsorbed  $\text{NH}_3$  species proved the deduction that only after the titanium site in  $\text{Fe}\text{--}\text{O}\text{--}\text{Ti}$  structure occupied by  $\text{NH}_3$  was liberated, bridging nitrate could form on catalyst surface.

Finally, the evolution of surface species over  $\text{FeTiO}_x$  catalyst in the real SCR condition was investigated (Fig. 4C). After  $\text{NH}_3 + \text{NO} + \text{O}_2$  introduction for 1 min, besides of the adsorbed  $\text{NH}_3$  species occurred on catalyst surface, a small amount of monodentate nitrate was also detected, which could further react with ionic  $\text{NH}_4^+$  to form  $\text{N}_2$ . From 5 to 60 min, no nitrate species was present on the surface and only adsorbed  $\text{NH}_3$  species were observed.  $\text{NH}_2$  species after H-abstraction of coordinated  $\text{NH}_3$  also showed up at  $1527\text{ cm}^{-1}$ . This result suggested that the reaction between  $\text{NH}_4^+$  and monodentate nitrate or between  $\text{NH}_2$  and  $\text{NO}$  was very fast.

## 4. Discussion

### 4.1. SCR mechanism in the low temperature range

$\text{NH}_3$ -TPD results in Section 3.1 combined with the *in situ* DRIFTS of  $\text{NH}_3$  desorption in Fig. S1 have already revealed that both Brønsted and Lewis acid sites existed on the surface of  $\text{FeTiO}_x$  catalyst. Long et al. [6,31] reported that there were a larger proportion of Brønsted acid sites than Lewis acid sites at low temperatures on  $\text{Fe}^{3+}\text{--TiO}_2\text{--PILC}$  catalyst, which is consistent with our results. As we described in Section 3.1,  $\text{NH}_3$  mainly adsorbed on titanium sites of  $\text{FeTiO}_x$  catalyst in the form of ionic  $\text{NH}_4^+$  and coordinated  $\text{NH}_3$ . However, the coordinated  $\text{NH}_3$  showed less reactivity in the low temperature range because of its higher activation energy to form  $\text{NH}_2$  species [32]. Therefore,  $\text{NH}_4^+$  was the main reactive adsorbed  $\text{NH}_3$  species in the SCR condition below  $200^\circ\text{C}$ .

Furthermore,  $\text{NO}$  was easy to be oxidized by  $\text{O}_2$  over  $\text{Fe}^{3+}$  [7] and then further adsorb onto the iron sites to form nitrate species (mainly monodentate nitrate in the SCR condition as we described in Sections 3.1 and 3.4). Over the  $\text{FeTiO}_x$  catalyst at low temperatures, the reactive monodentate nitrate on iron sites could react with two neighboring  $\text{NH}_4^+$  on titanium sites to form intermediate species, which could further react with gaseous or weakly adsorbed  $\text{NO}$  to form  $\text{N}_2$  [33,34]. Thereupon, we can propose an L–H mechanism dominant for the SCR reaction over  $\text{FeTiO}_x$  catalyst in the relatively low temperature range ( $<200^\circ\text{C}$ ), as shown in Fig. 5A. In this mechanism, gaseous  $\text{O}_2$  played an important role in  $\text{NO}$  oxidation. The overall reaction after the formation of intermediate species is as follows:  $2\text{NH}_4^+ + \text{NO}_2(\text{ads}) + \text{NO} \rightarrow 2\text{N}_2 + 3\text{H}_2\text{O} + 2\text{H}^+$ , which is exactly the “fast SCR” mechanism proposed by other researchers [35,36]. This explains the reason for the significant enhancement in the low temperature SCR activity over  $\text{FeTiO}_x$  catalyst when the  $\text{NO}_2/\text{NO}$  molar ratio was adjusted to 1:1 in our previous study [2]. Under the SCR condition, as the *in situ* DRIFTS results shown, only adsorbed  $\text{NH}_3$  species was detected on the surface, implying that once the active monodentate nitrate was formed, it could be consumed quickly to produce the final products. Therefore, the formation of monodentate nitrate on iron sites was actually the rate-determining step for the overall SCR reaction.



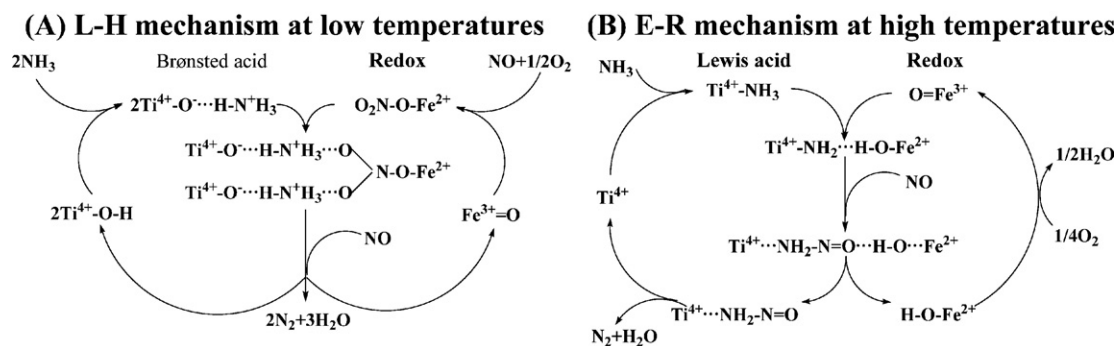


Fig. 5. Proposed SCR mechanisms over  $\text{FeTiO}_x$  catalyst in different temperature ranges.

#### 4.2. SCR mechanism in the high temperature range

At relatively high temperatures ( $>200^\circ\text{C}$ ), some Brønsted acid sites could be transformed into Lewis acid sites due to the dehydroxylation and dehydration effect [31]. As the *in situ* DRIFTS results shown in Section 3.4, the coordinated  $\text{NH}_3$  adsorbed on titanium sites could further undergo oxidative dehydrogenation by neighboring  $\text{Fe}^{3+}$  to form  $\text{NH}_2$  intermediate species. Moreover, as the TPSR results shown in Section 3.2,  $\text{NO}$  could react with  $\text{NH}_2$  species directly to form  $\text{NH}_2\text{NO}$  intermediate species without oxidation by  $\text{O}_2$ . Because of the low thermal stability and rapid decomposition of  $\text{NH}_2\text{NO}$  into  $\text{N}_2$  and  $\text{H}_2\text{O}$ , no infrared band attributed to  $\text{NH}_2\text{NO}$  was observed on catalyst surface under the SCR condition. Therefore, the formation of  $\text{NH}_2\text{NO}$  was actually the rate-determining step in the overall SCR reaction. During this process,  $\text{Fe}^{3+}$  was firstly reduced to  $\text{Fe}^{2+}$  in the H-abstraction step of coordinated  $\text{NH}_3$ , and then reoxidized to  $\text{Fe}^{3+}$  by  $\text{O}_2$  to complete a redox circle. Accordingly, we can propose an E–R mechanism for the SCR reaction over  $\text{FeTiO}_x$  catalyst as shown in Fig. 5B, which is mainly predominant in the high temperature range ( $>200^\circ\text{C}$ ). The similar SCR reaction pathway was also proposed on  $\text{V}_2\text{O}_5\text{--WO}_3/\text{TiO}_2$  [29],  $\text{MnO}_x\text{--CeO}_2$  [37] and  $\text{Fe}_2\text{O}_3\text{--WO}_3/\text{ZrO}_2$  [8] catalysts by other researchers.

#### 4.3. Enlightenment

The comprehensive understanding of  $\text{NH}_3\text{--SCR}$  mechanisms in different temperature ranges can supply theoretical guidance for the catalyst redesign and activity improvement for environmental-friendly  $\text{FeTiO}_x$  catalyst, even for other Fe-based SCR catalysts. To further enhance the low temperature  $\text{DeNO}_x$  efficiency, the L–H reaction pathway between ionic  $\text{NH}_4^+$  and monodentate nitrate at low temperatures should be strengthened. Therefore, for  $\text{FeTiO}_x$  SCR catalyst of which the  $\text{NH}_3$  adsorption ability is already strong enough, a wise method is to increase its  $\text{NO}$  oxidation ability and thus to increase the formation of reactive monodentate nitrate. A good example is to substitute partial Fe in  $\text{FeTiO}_x$  catalyst by Mn with higher oxidative ability, through which the formation of monodentate nitrate is largely enhanced and the low temperature  $\text{DeNO}_x$  efficiency is also obviously improved [38].

In our previous study, it has been concluded that even over the  $\text{SO}_2$ -poisoned  $\text{FeTiO}_x$  catalyst, the  $\text{NO}_x$  conversion still could be maintained above 90% from 250 to  $400^\circ\text{C}$  [1]. This is mainly owing to the mechanistic nature of E–R reaction pathway between  $\text{NH}_2$  species and  $\text{NO}$  at high temperatures, on which the deposited sulfate species showed no inhibition effect at all.

It is optimal for us to find a certain SCR catalyst, over which the  $\text{NH}_3\text{--SCR}$  reaction follows an E–R mechanism between  $\text{NH}_2$  and  $\text{NO}$  at low temperatures, or the sulfate species could decompose readily at low temperatures on the catalyst surface, thus we can obtain high  $\text{DeNO}_x$  efficiency and high  $\text{SO}_2$  durability simultaneously even

below  $100^\circ\text{C}$ . This can be a struggling aim for  $\text{DeNO}_x$  researchers, although there is still a long journey to conquer.

## 5. Conclusions

Both Brønsted acid sites and Lewis acid sites existed on the surface of iron titanate catalyst  $\text{FeTiO}_x$ . During the SCR process,  $\text{NH}_3$  mainly adsorbed onto the titanium sites in a form of ionic  $\text{NH}_4^+$  and coordinated  $\text{NH}_3$ , while  $\text{NO}_x$  mainly adsorbed onto the iron sites in a form of monodentate nitrate species.

In the low temperature range ( $<200^\circ\text{C}$ ), the reactive surface species were mainly  $\text{NH}_4^+$  and monodentate nitrate, and the SCR reaction mainly followed the L–H mechanism, during which the formation of monodentate nitrate species resulting from the  $\text{NO}$  oxidation by  $\text{O}_2$  over  $\text{Fe}^{3+}$  sites was the rate-determining step.

In the high temperature range ( $>200^\circ\text{C}$ ), the reactive surface species were mainly  $\text{NH}_2$ , and the SCR reaction mainly followed the E–R mechanism, during which the formation of  $\text{NH}_2\text{NO}$  intermediate species after H-abstraction of  $\text{NH}_3$  by neighboring  $\text{Fe}^{3+}$  sites was the rate-determining step.  $\text{O}_2$  took effect in the reoxidation of  $\text{Fe}^{2+}$  (produced in the H-abstraction step) to  $\text{Fe}^{3+}$  to complete a redox cycle.

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## Appendix A. Supplementary data

Supplementary data associated with this article can be found, in the online version, at doi:10.1016/j.cattod.2011.02.049.

## References

- [1] F. Liu, H. He, C. Zhang, Chem. Commun. (2008) 2043.
- [2] F. Liu, H. He, C. Zhang, Z. Feng, L. Zheng, Y. Xie, T. Hu, Appl. Catal. B: Environ. 96 (2010) 408.
- [3] F. Liu, H. He, J. Phys. Chem. C 114 (2010) 16929.
- [4] F. Liu, K. Asakura, H. He, Y. Liu, W. Shan, X. Shi, C. Zhang, Catal. Today 164 (2011) 488.
- [5] G. Busca, L. Lietti, G. Ramis, F. Berti, Appl. Catal. B: Environ. 18 (1998) 1.
- [6] R.Q. Long, R.T. Yang, J. Catal. 190 (2000) 22.
- [7] R.Q. Long, R.T. Yang, J. Catal. 207 (2002) 224.
- [8] N. Apostolescu, B. Geiger, K. Hizbullah, M.T. Jan, S. Kureti, D. Reichert, F. Schott, W. Weisweiler, Appl. Catal. B: Environ. 62 (2006) 104.
- [9] R.Q. Long, R.T. Yang, J. Am. Chem. Soc. 121 (1999) 5595.
- [10] R.Q. Long, R.T. Yang, J. Catal. 186 (1999) 254.
- [11] P. Balle, B. Geiger, S. Kureti, Appl. Catal. B: Environ. 85 (2009) 109.
- [12] M. Iwasaki, K. Yamazaki, K. Banno, H. Shinjoh, J. Catal. 260 (2008) 205.

- [13] J.-H. Park, H.J. Park, J.H. Baik, I.S. Nam, C.-H. Shin, J.-H. Lee, B.K. Cho, S.H. Oh, *J. Catal.* 240 (2006) 47.
- [14] L.S. Cheng, R.T. Yang, N. Chen, *J. Catal.* 164 (1996) 70.
- [15] R.Q. Long, R.T. Yang, *J. Catal.* 207 (2002) 158.
- [16] L. Chmielarz, R. Dziembaj, T. Grzybek, J. Klinik, T. Łojewski, D. Olszewska, A. Węgrzyn, *Catal. Lett.* 70 (2000) 51.
- [17] R.Q. Long, R.T. Yang, *J. Catal.* 198 (2001) 20.
- [18] Z. Liu, P.J. Millington, J.E. Bailie, R.R. Rajaram, J.A. Anderson, *Micropor. Mesopor. Mater.* 104 (2007) 159.
- [19] K.I. Shimizu, J. Shibata, A. Satsuma, *J. Catal.* 239 (2006) 402.
- [20] W.S. Kijlstra, D.S. Brands, H.I. Smit, E.K. Poels, A. Bliet, *J. Catal.* 171 (1997) 219.
- [21] Z. Liu, K.S. Oh, S.I. Woo, *Catal. Lett.* 120 (2008) 143.
- [22] G. Piazzesi, M. Elsener, O. Kröcher, A. Wokaun, *Appl. Catal. B: Environ.* 65 (2006) 169.
- [23] G.M. Underwood, T.M. Miller, V.H. Grassian, *J. Phys. Chem. A* 103 (1999) 6184.
- [24] D.A. Peña, B.S. Uphade, E.P. Reddy, P.G. Smirniotis, *J. Phys. Chem. B* 108 (2004) 9927.
- [25] H.Y. Chen, T. Voskoboinikov, W.M.H. Sachtler, *J. Catal.* 180 (1998) 171.
- [26] W.S. Kijlstra, D.S. Brands, E.K. Poels, A. Bliet, *J. Catal.* 171 (1997) 208.
- [27] G. Ramis, M.A. Larrubia, G. Busca, *Top. Catal.* 11–12 (2000) 161.
- [28] G. Busca, M.A. Larrubia, L. Arrighi, G. Ramis, *Catal. Today* 107–108 (2005) 139.
- [29] L. Lietti, G. Ramis, F. Berti, G. Toledo, D. Robba, G. Busca, P. Forzatti, *Catal. Today* 42 (1998) 101.
- [30] N.Y. Topsøe, *Science* 265 (1994) 1217.
- [31] R.Q. Long, M.T. Chang, R.T. Yang, *Appl. Catal. B: Environ.* 33 (2001) 97.
- [32] G. Qi, R.T. Yang, *J. Phys. Chem. B* 108 (2004) 15738.
- [33] V. Sanchez-Escribano, T. Montanari, G. Busca, *Appl. Catal. B: Environ.* 58 (2005) 19.
- [34] J. Li, R. Zhu, Y. Cheng, C.K. Lambert, R.T. Yang, *Environ. Sci. Technol.* 44 (2010) 1799.
- [35] A. Grossale, I. Nova, E. Tronconi, D. Chatterjee, M. Weibel, *J. Catal.* 256 (2008) 312.
- [36] M. Devadas, O. Kröcher, M. Elsener, A. Wokaun, N. Söger, M. Pfeifer, Y. Demel, L. Mussmann, *Appl. Catal. B: Environ.* 67 (2006) 187.
- [37] F. Eigenmann, M. Maciejewski, A. Baiker, *Appl. Catal. B: Environ.* 62 (2006) 311.
- [38] F. Liu, H. He, Y. Ding, C. Zhang, *Appl. Catal. B: Environ.* 93 (2009) 194.